Tolerance of piava juveniles to different ammonia concentrations

Tolerância de juvenis de piava para diferentes concentrações de amônia na água

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Abstract

The objective of this study was to evaluate the tolerance of piavas (Leporinus obtusidens) to ammonia (NH₃) by measuring its effects on plasma ion levels and tissue metabolic parameters. Piava juveniles (25-30 g) were exposed to five concentrations of NH₃ (mg L⁻¹): 0.003 (control), 0.1, 0.4, 0.7, and 1.4; after 96 hours, plasma levels of Na⁺, K⁺, Cl⁻ and NH₄⁺ ions and metabolic and enzyme activity in tissues (liver, kidneys, gills and muscle) were measured. The lethal concentration (LC50; 96h) of NH₃ was 0.27 mg L⁻¹. As NH₃ increased, Na⁺ and NH₄⁺ in the plasma increased and K⁺ decreased. In addition, Na⁺/K⁺-ATPase activity concomitantly increased in the gills and decreased in the kidneys. Glucose, glycogen, and protein levels decreased, while lactate and ammonia increased in the tissues of piava juveniles that were treated with higher concentrations of ammonia. The observed lethal toxicity could be due to a gradual depletion of plasma ion levels and a reduction of metabolic and Na⁺/K⁺-ATPase activity in tissues. Both can lead to dysfunction in ionoregulatory and physiological systems. This finding has implications for the management of fish culture of piavas.

Key words: Ionoregulatory, lethal concentration, Na⁺/K⁺-ATPase

Resumo

O objetivo deste trabalho foi avaliar a tolerância de piava (Leporinus obtusidens) a diferentes concentrações de amônia da água e os efeitos sobre seus íons plasmáticos e parâmetros metabólicos. Juvenis de piava (25-30 g) foram expostos a diferentes concentrações de amônia da água (mg/L NH₃): 0,003 (controle); 0,1; 0,4; 0,7 e 1,4, por 96 horas. Foram analisados níveis de Na⁺, K⁺, Cl⁻ e NH₄⁺ no plasma e metabolismo e atividade enzimática nos tecidos (fígado, rim, brânquias e músculo). A concentração letal (LC50; 96h) foi de 0,27 mg/L de NH₃, e os níveis plasmáticos de Na⁺ e NH₄⁺ aumentaram, enquanto os níveis de K⁺ diminuíram com o aumento das concentrações de NH₃ na água. A atividade de Na⁺/K⁺-ATPase aumentou nas brânquias e diminuiu nos rins com o aumento das concentrações de NH₃ na água. Níveis de glicose, glicogênio e proteína (exceto no músculo) também reduziram, mas níveis de lactato aumentaram nos tecidos dos juvenis expostos as maiores concentrações de NH₃ na água. A toxicidade letal observada poderia ser devido a depleção gradual de ions no plasma, a redução do metabolismo e a atividade nos tecidos de Na⁺/K⁺-ATPase. O aumento das concentrações de amônia na água leva ao acúmulo de amônia no plasma e nos tecidos, o que reduz o gradiente de difusão plasma-água de NH₃ e induz alterações metabólicas e ionoregulatórias nas piavas, o que tem implicações para o manejo de piavas.

Palavras-chave: Ionoregulação, concentração letal, Na⁺/K⁺-ATPase, manejo

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Introduction

Teleost fish are primarily ammoniotelic, and ammonia is the principal end-product of protein catabolism in most fish species. High ammonia concentrations may occur in intensive fish culture and can become unsafe unless ammonia is removed continuously either by biological filtration or periodical water change (HEGAZI et al., 2010). The toxicity of ammonia increases with water pH because its unionized form (NH₃) can diffuse across the gill membranes via the Rh glycoproteins (WRIGHT; WOOD, 2012), which then increases the proportion of NH₃ in the water and subsequently decreases the diffusion gradient (ERDOGAN et al., 2005).

Ammonia is toxic to fishes and can cause convulsions, coma, and death. These consequences most likely occur because the elevated levels of NH₄⁺ displace K⁺, which increases neuronal firing rate (ERDOGAN et al., 2005; WALSH et al., 2007). Chronic or sublethal exposure to NH₃ can also cause degeneration of tissues in the gills (MIRON et al., 2008). Sub-lethal ammonia concentrations decrease the growth rate of fish (MIRON et al., 2011; FERREIRA et al., 2013), interfering in the artificial fertilization (VIDAL et al., 2013) and may cause several physiological and histological changes (WRIGHT; WOOD, 2012; LIEW et al., 2013). For these reasons, toxicity of ammonia to fish has been intensively investigated in numerous fish species (BENLİ; KÖKSAL, 2005; ERDOGAN et al., 2005; MIRON et al., 2008; LIEW et al., 2013; FERREIRA et al., 2013).

Tolerance to high environmental ammonia concentrations varies among freshwater species (MIRON et al., 2008; LIEW et al., 2013). Species such as Opsanus beta, Opsanus tau, and Porichthys notatus are known to tolerate high concentrations of ammonia in the water (WANG; WALSH, 2000). Such highly tolerant species may have evolved specialized strategies that may reduce the toxicity of ammonia through physiological alterations (RANDALL; TSUI, 2002). For example, some species produce derived compounds, such as urea in Clarias batrachus (SAHA et al., 2002) and glutamine in Oncorhynchus mykiss (rainbow trout) (WICKS; RANDALL, 2002). Therefore, tolerance concentrations should be determined according to species-specific thresholds (SCHRAM et al., 2010).

Piava, Leporinus obtusidens Valenciennes (1847), can be found in Brazil, mainly in the São Francisco, Paraná and the Uruguai Basins and is an important native species for fish culture in South Brazil (COPATTI; AMARAL, 2009; REYNALTE-TATAJE; ZANIBONI-FILHO, 2010) that adapts well to different stocking density and feeding frequency (COPATTI et al., 2008). No study has analyzed the tolerance of piavas to the acute exposure to ammonia and the effects on their ionoregulatory and metabolic parameters. The objective of this work was to provide this information.

Materials and Methods

Piava juveniles (8-11 cm; 26-29 g) were purchased from a local fish culture, transported to the Fish Physiology Laboratory at the Universidade Federal de Santa Maria, and kept in continuously aerated 250 L tanks (23 °C, pH 7.4, water hardness 30 mg L⁻¹ CaCO₃) for two weeks. After acclimation (10 days), the fish were then transferred randomly into continuously aerated 40 L tanks (10 juveniles per tank) with the following NH₃ concentrations: 0.003 (control), 0.1 ± 0.06, 0.4 ± 0.11, 0.7 ± 0.13, and 1.4 ± 0.19 mg L⁻¹ (in triplicate). The concentrations were chosen according to previous studies of Serafini et al. (2009) for dourado (Salminus brasiliensis) and Ferreira et al. (2013) for silver catfish (Rhamdia quelen). The fish were first kept in their respective tanks for 96 h, and then NH₄Cl was added to increase NH₃ to experimental concentrations.

The tanks were siphoned daily to remove residues and feces, and at least 20-40% of the water was replaced with water that had been adjusted according to the experimental NH₃ level. The fish...
were not fed during the experiment, the tanks were checked for dead fish every two hours, and the number of dead fish was recorded.

For all individuals, tissues from the liver, kidneys, muscle, and gills were collected at the end of the experiment or when fish showed complete equilibrium loss and very low opercular movements (which tended to occur at the higher NH$_3$ concentrations). Specimens were stored in liquid nitrogen until analyses. Tissue samples were homogenized in 100 mg L$^{-1}$ 20% trichloride acetic acid using a Potter-Elvehjem type homogenizer (1000 rpm per 3 min in ice bath) and then centrifuged at 3000 x g for 5 min. The supernatants were used to determine the concentrations of ammonia (VERDOUW et al., 1978), glucose, glycogen (DUBOIS et al., 1956), lactate (HARROWER; BROWN, 1972), and protein (LOWRY et al., 1951) in the liver, kidneys and muscle tissues.

To assay Na$^+$/K$^+$-ATPase in gills and kidneys, the tissue samples were homogenized with a Potter-Elvehjem type homogenizer with a 1:5 ratio of SET (sacarose-EDTA-TRIS; pH 7.5) buffer. Next, the homogenate was centrifuged for 5 min at 6000 x g, and the supernatant was used for enzyme assays adapted from (FLIK et al., 1983). Blood was collected from the caudal vein with heparinized 1 mL syringes and centrifuged at 1200 x g for 5 min to separate the plasma. The Na$^+$ and K$^+$ levels were analyzed with a B262 flame spectrophotometer (Micronal, São Paulo, Brazil); the Cl$^-$ levels were determined according to the protocol described by Zall et al. (1956).

The water quality parameters were kept constant to avoid variation in NH$_3$ toxicity. The dissolved oxygen levels and temperature were measured every two hours using an YSI oxygen meter (model Y5512, Yellow Springs, USA), and temperature was maintained at 25 °C using an air conditioner in the laboratory. Water pH was monitored with a BNC Oakton pH meter (model 400A) and sustained within a neutral range (7.0-7.2). Water samples were collected every two hours and frozen for analysis of the total ammonia as described by Verdouw et al. (1978); NH$_3$ concentrations were calculated according to Colt (2002). Water hardness and alkalinity were determined using the EDTA titrimetric method (EATON et al., 2005). At the time of water sample collection, the behavior of the fish (e.g., swimming and equilibrium) was observed.

The data are reported as the mean ± SEM and 95% confidence intervals. Homogeneity of variances among the groups was determined with a Levene test. The data showed homogeneous variances. The lethal concentration of NH$_3$ at 96 h (LC50; 96 h) was calculated by the probit method (FINNEY, 1971), an analysis which has been widely used for many years to calculate LC50. The relationships between NH$_3$ concentrations and metabolic parameters were examined using Sigma Plot 8.0 software. If no significant relationship was found, comparisons between treatments were performed with one-way analysis of variance and Tukey test using Statistica software (version 7) (P < 0.05).

Results

Only the control (0.03 mg L$^{-1}$) and 0.1 mg L$^{-1}$ NH$_3$ treatment groups exhibited 100% survival. The exposure of piava to NH$_3$ concentrations of 0.4 mg L$^{-1}$ or higher caused altered locomotive behavior and then death. Mortality was 78%, 100% (after 72 h), and 100% (after 48 h) at 0.4, 0.7, and 1.4 mg L$^{-1}$ NH$_3$, respectively. The LC$_{50-96h}$ of NH$_3$ for piava is 0.27 mg L$^{-1}$ (confidence interval: 0.21-0.34 mg L$^{-1}$). Plasma levels of Na$^+$ and NH$_4^+$ increased with NH$_3$, whereas levels of K$^+$ decreased and Cl$^-$ remained stable (Figure 1).

Total NH$_3$ increased in the liver, kidneys, and muscle and Na$^+$/K$^+$-ATPase activity increased in the gills and decreased in the kidneys (Figures 2 and 3). The lactate levels in the liver, kidneys, and muscle increased with the NH$_3$ concentrations (Figure 4). In contrast, glucose, glycogen, and protein decreased in the liver and kidneys, and glucose and glycogen decreased in muscle (Figures 5, 6 and 7).
Figure 1. Ion and NH$_3^+$ plasma levels in piava (y) as a function of NH$_3$ concentrations (x). The following equations were fitted to the data: A= Na': $r^2 = 0.92$, $y = 84.17 + 139.14x$; B= Cl': no relationship or significant difference; C= K': $r^2 = 0.97$, $y = 8.02 - 3.82x$; D= NH$_4^+$: $r^2 = 0.94$, $y = 8.31 + 14.08x$. 
Figure 2. Ammonia levels in the tissues of piava (y) as a function of NH3 concentrations (x). The following equations were fitted to the data: A= liver: \( r^2 = 0.89 \), \( y = 89.12 + 165.01x \); B= kidneys: \( r^2 = 0.96 \), \( y = 57.24 + 84.98x \); C= muscle: \( r^2 = 0.95 \), \( y = 104.58 + 194.23x \).

Figure 3. Na⁺/K⁺-ATPase activity in the gills and kidneys of piava (y) as a function of NH3 concentrations (x). The following equations were fitted to the data: A= gills: \( r^2 = 0.93 \), \( y = 14.38 + 42.13x \); B= kidneys: \( r^2 = 0.88 \), \( y = 31.68 - 18.10x \).
Figure 4. Lactate levels in the tissues of piava (y) as a function of NH₃ concentrations (x). The following equations were fitted to the data: A= liver: r² = 0.97, y = 86.41 + 72.13x; B= kidneys: r² = 0.87, y = 74.35 + 57.82x; C= muscle: r² = 0.93, y = 1382.24 + 1133.93x.

Figure 5. Glucose levels in the tissues of piava (y) as a function of NH₃ concentrations (x). The following equations were fitted to the data: A= liver: r² = 0.91, y = 43.11 – 22.49x; B= kidneys: r² = 0.85, y = 12.37 – 7.09x; C= muscle: r² = 0.89, y = 65.94 – 43.50x.
Figure 6. Glycogen levels in the tissues of piava (y) as a function of NH₃ concentrations (x). The following equations were fitted to the data: A= liver: $r^2 = 0.97$, $y = 109.19 - 41.85x$; B= kidneys: $r^2 = 0.93$, $y = 71.49 - 31.30x$; C= muscle: $r^2 = 0.86$, $y = 128.67 - 69.58$.

Figure 7. Protein levels in the tissues of piava (y) as a function of NH₃ concentrations (x). The following equations were fitted to the data: A= liver: $r^2 = 0.98$, $y = 80.48 - 33.49x$; B= kidneys: $r^2 = 0.90$, $y = 51.19 - 29.34x$; C= muscle: no relationship or significant difference.
Discussion

In this experiment, piavas that were exposed to the higher concentrations of NH$_3$ demonstrated responses that are typical of fish that are under short-term exposure to ammonia, including loss of equilibrium, erratic swimming, and death (RANDALL; TSUI, 2002; WALSH et al., 2007). The observed behaviors are consistent with previous descriptions of the damaging effects on the central nervous system that may be caused by acute NH$_3$ toxicity (RANDALL; TSUI, 2002; WALSH et al., 2007). Exposure to sub-lethal concentrations of NH$_3$ halts the metabolism of amino acids and mitochondria in the brain (WALSH et al., 2007). Similar results were described for sea bass (Dicentrarchus labrax), sea bream (Sparus aurata), and turbot (Scophthalmus maximus) juveniles that were exposed to 6.43, 4.87 and 5.70 mg L$^{-1}$ NH$_3$, respectively, for 96 h (RUYET et al., 1995). Seabass that were exposed to different NH$_3$ concentrations (0.71, 0.88 and 0.90 mg L$^{-1}$) for 63 days showed a positive relationship between plasma ammonia and NH$_3$ concentrations (LEMARIÉ et al., 2004). The high NH$_3$ concentrations may have also affected ammonia levels of other tissues (Fig. 2).

After exposure to 0.11 mg L$^{-1}$ NH$_3$ for 12 h and up to seven days, Na$^+$ uptake increased in rainbow trout, goldfish (Carassius auratus), and carp (Cyprinus carpio), possibly due to the activation of the branchial apical “Na$^+$/NH$_4^+$ exchange metabolon” (LIEW et al., 2013). It is likely that this mechanism was also activated in the piava with the replacement of NH$_4^+$ by K$^+$ in the basolateral Na$^+$/K$^+$/ATPase, which led to an increase in Na$^+$ and a decrease in K$^+$ plasma levels.

Liew et al. (2013) observed that diffusive water efflux rates and net K$^+$ loss rates across the gills were enhanced during exposure to 0.11 mg L$^{-1}$ NH$_3$ in rainbow trout, which indicates increased gill transcellular permeability. The net K$^+$ loss rates across the gills must have occurred in this study, since plasma levels of K decreased at the highest NH$_3$ concentrations. On the other hand, silver catfish exposed to 0.1 mg L$^{-1}$ NH$_3$ for 6-24 h had higher plasma Na$^+$, K$^+$ and Cl$^-$ levels compared to those exposed to 0.03 mg L$^{-1}$ NH$_3$ (BECKER et al., 2009). The effects of NH$_3$ on plasma ion levels of freshwater fish seem to be highly species-dependent and vary according to the pH and NH$_3$ concentration of the water (BOLNER et al., 2014).

The low resistance of piava to ammonia toxicity explains the physiological changes described in this study. In this study, plasma ammonia levels were expected to increase with NH$_3$ concentration because higher NH$_3$ concentrations would decrease the plasma-water gradient for NH$_3$ diffusion. As a consequence, NH$_3$ excretion would decrease. Seabass, seabream, and turbot juveniles that were exposed to 6.43, 4.87, and 5.70 mg L$^{-1}$ NH$_3$, respectively, for 96 h also exhibited increased plasma concentrations of NH$_3$ (RUYET et al., 1995). Seabass that were exposed to different NH$_3$ concentrations (0.71, 0.88 and 0.90 mg L$^{-1}$) for 63 days showed a positive relationship between plasma ammonia and NH$_3$ concentrations (LEMARIÉ et al., 2004). The high NH$_3$ concentrations may have also affected ammonia levels of other tissues (Fig. 2).
At high NH₃ concentrations, Na⁺/K⁺-ATPase activity increased in the gills but decreased in the kidneys of piavas. This finding was expected because upregulation of the “Na⁺/NH₄⁺ exchange metabolon” in the gills is often associated with increased gene expression and/or enzyme activity of Na⁺/K⁺-ATPase (WRIGHT; WOOD, 2012) in order to efficiently eliminate the excess ammonia. Silver perch (*Bidyanus bidyanus*) and gold perch (*Macquaria ambigua*) exposed to 5 mg L⁻¹ NH₃ for 120 h also displayed increased Na⁺/K⁺-ATPase activity in the gills (ALAM; FRANKEL, 2006). In freshwater fish, kidneys produce a very dilute urine (BOLNER; BALDISSEROTTO, 2007), and the reduction in Na⁺/K⁺-ATPase activity at high NH₃ concentrations suggests a deficiency in tubular ion reabsorption.

Stressed fish are expected to increase plasma glucose and decrease glycogen levels to contribute to the increase in energy demand (BARCELLOS et al., 2004). However, piava exposed to high NH₃ concentrations (>).0.4 mg L⁻¹), had permanent stress, demonstrated by the higher mortality rates and lower glucose, glycogen and protein levels than control group and 0.1 mg L⁻¹ NH₃ treatment after 96 h. Silver catfish exposed to 0.50 mg L⁻¹ NH₃ increased significantly plasma glucose in the first 24 h, but presented 100% mortality within 5 days (BALDISSEROTTO et al., 2014).

So, the reduction in glucose and glycogen in the tissues can be explained by the use of these substrates in the energetic demand required for NH₃ excretion. The stress of the exposure to high NH₃ concentrations results in an increase of plasma cortisol levels in fish (LIEW et al., 2013), which stimulates glycolysis and gluconeogenesis and increases protein catabolism (RANDALL; TSUI, 2002), possibly due to catecholamine-mediated glycolysis and cortisol-mediated gluconeogenesis (PANKHURST, 2011). This process likely occurred in piava that were exposed to high NH₃ concentrations, resulting in lower protein levels in the tissues due to the catabolic process. The metabolism of piava juveniles could be an adaptation that has evolved in response to the energetic demands of NH₃ excretion.

Lactate is a final product of anaerobic glycolysis during intense exercise or stress, due to poor muscle oxygenation or alteration of the aerobic respiration (WENDELAAR BONGA, 1997). In the present study, the increase in lactate in the tissues suggests that the increase in NH₃ concentrations caused the piava to employ an anaerobic route with the use of lactate as substrate for gluconeogenesis. On the other hand, Baldisserotto et al. (2014) did not show alterations in plasma lactate levels in silver catfish exposed to 0.18 or 0.50 mg L⁻¹ NH₃ on the first 24 h of experiment.

The results suggest that the observed lethal toxicity (0.27 mg L⁻¹ NH₃) could be related to a gradual disruption of ionic levels in the plasma and metabolic and Na⁺/K⁺-ATPase activity in tissues, both of which could then lead to the depression of physiological and ionoregulatory systems. Treatment with the highest concentrations of NH₃ (>0.4 mg L⁻¹) showed higher plasma levels of lactate and the lower glucose and glycogen levels. NH₃ is a stressor for piava, and its effect is immediate but concentration dependent. The results are helpful to fish culture farmers in preventing the decreased productivity that may be caused by elevated NH₃ concentrations (>0.1 mg L⁻¹ NH₃).

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References


